

Is the evolving sport of mountain biking compatible with fauna conservation in national parks?

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ABSTRACT

Historically, most people have tended to visit national parks for 'rest, relaxation and reinvigoration', typically resulting in moderate ecological impacts. However, increasingly, recreation in natural areas is including 'adventure' sports. One such recreation/sport that now incorporates a range of forms, including adventure derivatives, is mountain biking. In the more extreme forms, riders use extensive trails, often with steep segments and natural or human-made obstacles demonstrate technical skills (e.g., balance, calculated risk-taking, excitement, speed). Appreciation of the natural environment is seldom, if ever, a reason for participation. In this paper we consider the potential for impact on the fauna of national parks. While there is a dearth of information on the impact of mountain biking, we conclude that park management needs to be strategic in their consideration of the issues associated with mountain biking or the outcome will be further degradation of natural areas and, at the least, loss of many animals if not major threats to populations.

Key words: fauna management, adventure sports, native animals, environmental impact, bike trails, natural area, visitor attitudes

Introduction

Historically, most people have tended to visit national parks for 'rest, relaxation and reinvigoration', and their ecological impact was typically considered to be low (Hall *et al.* 2010). More recently; however, there has been increasing use of national parks for more active recreation, 'adventure' sports, including rock climbing, abseiling, canyoning, whitewater kayaking, skiing, off-road driving and mountain biking (Hardiman and Burgin, 2011a).

'Adventure recreation' is defined as 'outdoor activities in which the uncontrollable hazards of a natural environment or feature are deliberately challenged through the application of specially-developed skills and judgment' (Brown 1989, pp. 37). Such sports have been criticised for their impacts, for example rock bolting (Jones, 2004), vegetation loss (Groom *et al.* 2007) and soil erosion (Ewert and Hollenhorst 1994). However, more recently some forms of 'adventure recreation' have been morphing into 'extreme sports'. These comprise a constantly-evolving collection of new sports or extension of existing ones which pose an increased risk of ecological impacts. As the term implies, extreme sports typically push the existing boundaries of risk for thrill's own sake with the aim of inducing an adrenaline 'buzz' by overcoming fear. This is induced by speed, gravity, and/or height (Ewert *et al.* 2006, Carnicelli-Filho *et al.* 2010). Typical examples of extreme sports are speed rock climbing, waterfall kayaking, BASE (Buildings, Antennas, Spans, Earth) jumping, heliskiing, enduro motocross and downhill

mountain biking. Such forms of recreation typically depend on large, undeveloped landscapes, for example national parks or other protected areas with a biological conservation mandate (Ewert *et al.* 2006).

One difference between adventure recreation and extreme sports is the emphasis on thrill as an end product (Puchan 2005). Another difference is that extreme sports typically involve competition among participants. This further changes the mental dynamic and the motivation for visiting the natural area. Such aspects switch the primary objective from 'experiencing' (passive recreation) to 'conquering or beating' nature (Baker and Simon 2002). Extreme sports often commence as niche activities, either as a completely new derivative or from another, less extreme form and subsequently develop into mainstream forms of recreation in their own right, even becoming Olympic events (e.g., Starr *et al.* 2006; IFSC 2010). One example of an adventure recreational activity that is evolving into a more extreme version is mountain biking. With its increasing international popularity (Bradshaw 2006; Koepke 2005; Leberman and Mason 2000), there is an associated rising demand for purpose-built trails and infrastructure, potentially within national parks and other natural areas. In this paper we consider the evolving sport of mountain biking: is it compatible with nature conservation in national parks ~~or could it become an ecological disaster?~~

The evolving sport of mountain biking

Mountain biking is assumed to have originated in the United States of America (US) in the 1970s (White *et al.* 2006). It now has global participation (Koepke 2005), and has become probably the most popular land-based recreation in the world (White *et al.* 2006). Koepke (2005) estimated that between 1987 and 2000 the popularity of mountain biking increased four-fold. By 2003 approximately 10 million in the United States of America (US) regularly used mountain bikes (Green 2003) and approximately 21% of the US population is estimated to cycle on backcountry roads, trails, or across country at least annually (White *et al.* 2006). Although sales have slowed in recent years, mountain/hybrid bikes still represented 44% of unit sales through US speciality bicycle retailers between 2006-2008 (NBDA 2010). The sport is also increasingly popular in Europe. For example, in the United Kingdom (UK), 5.7% of the population is estimated to mountain bike regularly (TRC 2005), while elsewhere, for example in Germany (3.5 million mountain bikers), Switzerland and Austria (800,000 combined) participation is also growing (Koepke 2005). In Australia, the sport grew by 15.3% between 2001 and 2004 (Faulks *et al.* 2008), and of the 753,843 bikes sold in 2004, 70% were mountain bikes (Bradshaw 2006). Globally, the sport continues to grow in popularity, and the International Mountain Biking Association (IMBA) is today represented in 17 countries (IMBA 2010).

In the traditional form of mountain biking, i.e. 'cross-country/recreational', riders use lightweight bicycles to traverse a range of landscapes on rides typically lasting a few hours. Emphasis is on relaxation, exercise and appreciation of natural scenery (Chiu & Kriwoken 2003) especially on single-track trails where riders are segregated from cars and can enjoy a closer connection to nature (Koepke 2005). Although cross-country/recreational riding is still the most popular, mainstream form of the sport (89% of US; 97% of UK mountain bike riders), more physically challenging, extreme derivatives are growing in popularity. Such derivatives include 'downhill' (18% US; 22.2% UK), 'freeriding' (23% US; 21.2% UK) and 'trials' (14% US; 8.1% UK) (Cessford 1995; Green 2003; Koepke 2005; Gajda 2008; IMBA 2010).

'Downhill' mountain biking involves steep descent over rough terrain at high speed using heavy, specialist bikes with long-travel suspension. The focus is on speed, strength and calculated risk-taking to make the fastest possible descent against the clock. For protection, riders wear plastic body armour and full-face helmets. They are typically transported to the top of the racecourses/runs by ski lift, 4WD vehicle or helicopter (Cessford 1995; TRC 2005). In another extreme derivative, 'Free-riding' (*cf.* 'North Shore'), riders are challenged to overcome natural and/or man-made obstacles (e.g., elevated, narrow wooden boardways, log rides, ladder bridges and teeter-totters) on purpose-built circuits (Cessford 1995; Koepke 2005). The focus is on demonstrating technical skills, balance, calculated risk-taking, excitement, and speed. Similar to downhill, appreciation of the natural

environment is seldom, if ever, a reason for participation (see e.g., Tiedeman 2002; Leberman and Mason 2000; Chiu and Kriwoken 2003; Cessford 1995).

The increasing popularity and associated economic potential for 'destination mountain biking tourism' has been widely acknowledged. For example, US locations such as Moab (Utah) and Fruita (Colorado) each offer hundreds of kilometres of single track bike trails (MATC 2010; OTES 2010). Alpine ski resorts of Canada, such as Whistler Blackcomb, have more than 200 km of biking trails which include 34 trails of lift-serviced downhill routes (TRC 2005; Whistler Blackcomb undated). In Scotland, the 7stanes, part of the UK's 40-site network of dedicated mountain bike centres, offers 600 km of forest trails (TRC 2005). New Zealand also has extensive mountain biking opportunities (Cessford 1995; Ride Rotorua undated), and the sport is becoming increasingly popular in Australia (Chiu and Kriwoken 2003). This is reflected, for example in the expansion of mountain biking within Sydney (e.g., HSMBA undated).

Physical impacts of mountain biking

The rapidly increasing popularity of mountain biking, together with its concurrent evolution into different forms, has caused concern surrounding its potential ecological impacts. Ecological impacts associated with recreational trails generally emanate from their initial design and construction and subsequent use (e.g., type, user behaviour, frequency, and intensity; Sun and Walsh 1998). Problems in assessing such impacts are complex. For example, most natural area trail use is a shared resource with other forms of recreation, typically including bush walking, horse riding, and 4WD driving. The specific impacts due to mountain biking therefore often cannot be readily distinguished (Hendricks *et al.* 2001). Despite this, instances of the creation of unauthorised, informal bike trails and/or construction of technical track features such as concrete-reinforced jumps and wooden boardways used in freeriding/North Shore mountain biking are becoming commonplace in national parks in Australia (e.g., Davies and Newsome 2009).

On flat terrain under dry conditions, impacts on trails caused by recreational mountain biking, include increased water runoff and sediment yield, vegetation and species loss, and/or soil exposure generally have been found to be comparable with those of walking tracks, although less than motorised vehicle use or horse riding (Wilson and Seney 1994; Thurston and Reader 2001; Chiu and Kriwoken 2003). The potential for trail erosion, compaction, incision and widening from mountain biking is, however dependent on climate, slope and other environmental variables. Steep slopes with sparse vegetation and/or fine homogenous soils are most susceptible to damage (Goefit and Alder 2001; White *et al.* 2006). The greatest impacts from biking typically occur early in trail use, on downhill (braking and skidding) and uphill (wheel spinning) slopes (especially when wet), and on curves (braking and skidding) (Goefit and Alder 2001; White *et al.* 2006; Chiu and Kriwoken 2003). Such damage may increase trail incision, soil erosion, water runoff, and widening. However, limited research has

been undertaken on how such impacts may differ with use intensity (e.g., under racing conditions) or duration. In some locations, recent growth in mountain biking has meant that sites with previously only seasonal recreation are now exposed to year-round impacts. For example, in alpine ski resorts such as Canada's Whistler Blackcomb, summer revenue from mountain biking now represents approximately 75% of winter snow recreation revenue (TRC 2005, Whistler Blackcomb undated) eliminating the period for environmental recovery between ski seasons. Chiu and Kriwoken (2003) have shown that such impacts on erosion are cumulative, although curvilinear. After rapid initial erosion, the rate of change declines. The longitudinal studies needed to determine the long term chronic impacts of mountain biking are lacking, yet Buckley *et al.* (1999) noted, that even passive tourism causes impacts for the lifetime that the recreation activity is practised.

Although there is a paucity of comparative studies, the impacts on the flora and fauna of competitive mountain biking are likely to be greater than recreational biking and/or bush walking. This is because the essential thrill element of racing demands technically challenging courses involving steeper up/downhill slopes, faster, harder braking, more intense use, cutting corners, wet sections and jumps/drop offs, together with substantial vegetation trampling from riders and spectators off-track. A German study of a competitive mountain bike racing event showed soil compaction resulting from bikes occurred to a shallower depth compared to the impact of spectators. Compaction from the wheels of the bikes was less, but deeper and recovered within 19 months whereas the impacts of spectators persisted for longer (Wöhrstein 1998). Australian studies of such racing events have found that soil loss at sharp corners was greater than on straight sections (Hawes 1997). Under wetter conditions there were increased off-trail vegetation impacts and trail widening, especially on steep slopes and on corners. Racing under such conditions has also been shown to increase off-trail vegetation impacts and trail widening (Goefl and Alder 2001), although in another Australian study the damage reported was less severe (Chui and Kriwoken 2003).

Owing to the lack of comprehensive assessment of its impacts, especially over long term use (White *et al.* 2006), mountain biking remains restricted and/or banned in some ecologically fragile areas, such as parts of the Cairngorm Mountains in Scotland (Hanley *et al.* 2002) and wilderness zones of the Greater Blue Mountains World Heritage Area, Australia (NPWS 2001).

The impacts emanating from mountain biking are not necessarily unique; effectively all outdoor recreational pursuits in natural areas can have adverse effects on the local environment (Lynn and Brown 2003). However, it is the markedly different motivation driving participation, especially in the more extreme versions (e.g., excitement, risk, speed, competition - Tiedeman 2002; Leberman and Mason 2000; Chiu and Kriwoken 2003; Cessford 1995) and the construction of trails and/or infrastructure to gain such outcomes (Cessford 1995; TRC 2005), together with their extent (e.g., MATC 2010) and the very large number of participants involved that potentially sets this

sport apart. Unlike other forms of adventure recreation, which are typically niche activities, mountain biking, especially in its cross-country/recreational form, is today a true 'mass market' form of adventure recreation for all the family (TRC 2005; TRC/EKOS 2007).

Potential impacts on native fauna

Impacts of recreation, including mountain biking, on the physical environment (Priskin 2003; Chiu and Kriwoken 2003; Ewert *et al.* 2006; Hawes 1997) and associated flora (Whinam and Chilcott 1999; Groom *et al.* 2007; Pickering and Hill 2007) are typically the most obvious impacts in natural areas and the most commented upon (e.g., Symmonds *et al.* 2000; Leung and Marion 1996; Thurston and Reader 2001; Martin *et al.* 1989). In contrast, knowledge of recreational impacts on fauna is relatively limited (Taylor and Knight 2003; Knight and Cole 1995), especially regarding adventure recreation (Hardiman and Burgin, 2010a, 2011b, c), except where the species is the target of the recreational activity (e.g., bottlenose dolphin *Tursiops* sp., Shark Bay - Bejder *et al.* 2006; Mann *et al.* 2000; yellow eyed penguin *Megadyptes antipodes* - McClung *et al.* 2004; harp seals *Phoca groenlandica* - Kovacs and Innes 1990).

The immediate response of individual animals to recreational disturbance is usually either death or behavioural change (e.g., physiological, food habits, nest abandonment; Knight and Cole 1995; Lathrop 2003). Many of these behavioural responses are short term (Cassirer *et al.* 1992). For example, in a study that compared the effect (alert distance, flight distance, distance moved) of mountain bikers and hikers on bison *Bison bison*, mule deer *Odocoileus hemionus*, and pronghorn antelope *Antilocapra americana*, Taylor and Knight (2003) found no difference among species in response to the two types of recreationists. There was a 70% or greater chance (mule deer 96%) of recreationists flushing all three species from within 100 m from the trail. Birds have also been found to be sensitive to pedestrian traffic. For example, the black-crowned night heron *Nycticorax nycticorax* was observed to change behaviour in the presence of pedestrians (and canoeists), at least in the breeding season. In the presence of humans, birds spent more time scanning (increased vigilance), freezing (anti-predator behaviour) and less time grooming and sleeping (Esteban *et al.* 2007). Most such studies of disturbance of birds by pedestrian recreationists have found that there was a reduction in the size of the breeding population, presumably due to abandonment of the site, although the success of the remaining birds was typically not affected (Knight and Cole 1995). However, responses may vary. For example, Miller *et al.* (1998) studied the influence of recreational trails on bird breeding and found that composition of bird species was lower near trails than away from them and generalist species were more abundant near trails than specialists. Nest predation, however, was greater near trails than away from them. In contrast, Stake (2000) found no difference in density, return rates or age structure for the endangered golden-cheeked warbler *Dendroica chrysoparia* due to the introduction of mountain biking.

While limited data are available on the impacts of birds and larger vertebrates, publications were found that directly addressed the impact of mountain biking on small animals. However, such species (e.g., small mammals, reptiles, invertebrates) would be less able to avoid the approach of mountain bikers, who travel more quickly and quietly than walkers. Lathrop (2003) reported that the effects of direct mortality due to mountain biking were 'virtually' unstudied although he suggested that anecdotal evidence indicated that small mammals are vulnerable and are killed. In a study of 1976 members of Bicycle Queensland, Heesch *et al.* (2010) examined the correlates of cycling injury (demographic characteristics, reasons for cycling, years of cycling as an adult, cycling frequency) for the previous year. Approximately 10% were caused by a 'crash with an object on the road or path'. Since this included 'pothole, kerb, animal', and 'wet or gravel surface', we assumed that few accidents are reported to be caused by animals (including domestic animals). Despite these few recordings, moribund lizards and associated bicycle skid marks have been observed on an urban Brisbane bicycle track. It is likely that the most dangerous time for the local blue tongue lizard *Tiliqua scincoides* is soon after sunrise when local biking traffic is substantial and lizards take their first bask for the day. Snakes are also vulnerable: Australian red-bellied black snakes *Pseudechis porphyriacus* often lying across a track in the Blue Mountains National Park and are prone to being accidentally ridden over and killed. We predict that the risk of injury to animals is positively correlated with increasing numbers of bike riders in the same way that animal road kill is associated with high vehicle traffic levels.

Apart from direct injury by bikers, mountain biking trails (or any other pathway through natural areas) may indirectly impact on native species. Many small species are influenced by changes in the vegetation structure that occurs with disturbance at the edge of bushland. For example, the abundance of brown antechinus *Antechinus stuartii*, a small carnivorous marsupial, was found to respond to structural components of its habitat including understorey height and complexity, litter depth and the absence of logs (Knight and Fox 2000) all of which may be modified in the presence of a mountain biking track or built infrastructure (e.g. jumps, boardways, teeter-totters) in a national park.

Predation may also be greater at the interface of the tracks and natural areas. For example, Anderson and Burgin (2002) found that abundance of the small common *Lampropholis* skinks (*L. delicata*, *L. guichenoti*) at the edge of remnant bushland plots, separated only by the width of power line corridors, was only half that of sites located at the core of such small remnants of natural bushland, showing the detrimental 'edge effect' of the dividing corridors. In a later study, Anderson and Burgin (2008) provided evidence that these differences were sustained, and that bird predation was the major factor for the differences in abundance between the edge and core. Mountain biking trails through natural bushland offer an equivalent interface that has the potential to attract

animals, particularly reptiles that thermoregulate and expose them to predation and collision with bikers' wheels. For example, over a seven year period, Wotherspoon and Burgin (2011) collected 19 reptile species (33% of the local recorded reptile fauna) as road kill on early morning excursions in Faulconbridge on a suburban road in a 50 km zone that abutted national park. These species are also likely to access trails within the local national park and therefore expose themselves to possible collision if it were a mountain biking trail. Use of the trail would also potentially expose native animals to predators, including feral species such as the red fox *Vulpes vulpes* that penetrate natural areas by moving along such paths (Catling and Burt 1995).

Apart from their potential to act as a conduit for species to penetrate into natural areas (e.g., foxes – Catling and Burt 1995; toads *Rhinella marina* [*Bufo marinus*] – Seabrook and Dettman 1996; Brown *et al.* 2006), there is substantial evidence that roads or trails may act as barriers to the movement of animals due to behavioural avoidance, the presence of a physical barrier or development of a home range along the physical barrier (Donaldson and Bennett 2004). The extent to which roads act as barriers to dispersal depends on the physical characteristics of the road (e.g., clearing width, road surface, traffic density) and the characteristics of the species (e.g., species, size, mobility, habitat requirements). Small species for example beetles, spiders and snails with relatively limited mobility will be more affected than larger, more mobile species. The barriers may be physical, behavioural or sociological (Mader 1984; Baur and Baur 1990). Development of a home range along physical barriers has the potential to interfere with social interactions (Barnett *et al.* 1978; Burnett 1992). For example, after initial research into areas near roads, the eastern chipmunk *Tamias striatus*, was found to avoid roads and their verges. This avoidance was independent of traffic volume (Ford and Fahrig 2008). In contrast, a study of road kill in peri-urban Sydney and regional New South Wales, Burgin and Brainwood (2008) found that there were higher numbers of animals killed on medium volume traffic roads compared to lower volume, local traffic ones or major highways. They also observed that compared to where there were either barriers on both sides of the road or none present there were fewer road kills than when a barrier was present along one verge.

In the terrestrial environment, movement of smaller animals such as beetles and snails are presumably more greatly affected by such barriers (Baur and Baur 1990; Mader 1984) although they may increase their longitudinal movement parallel with the road. However, while roads inhibit movement of small mammals they rarely prevent movement across them (e.g., Oxley *et al.* 1974; Barnett *et al.* 1978). Since mountain bike trails tend to be narrow they would be less of a challenge than roads for most vertebrates. Reptiles may be the exception since they seek open areas and/or warm substrates as basking sites. For example, Wotherspoon and Burgin (2011) found that species considered locally rare were found in disproportionately higher numbers as road kills. Individuals of two species, eastern small-eyed *Cryptophis*

nigrescens and blind snake *Ramphotyphlops nigrescens* that had been seldom observed locally, were among the most common encountered road kills. While the low speed, suburban street that Wotherspoon and Burgin (2011) reported on has remained effectively unchanged for more than 20 years, species may become locally extinct as a result of new road development (Lunney *et al.* 2002) while the long-term viability of some vertebrate populations may be compromised (e.g., Jones 2000; Ramp and Ben-Ami 2006).

In addition to the age of the road, soils and/or habitat may also influence the impact on local species. For example, although mountain biking was not specifically mentioned, Ross *et al.* (2009) reported that bicycles contributed to the degradation of saltmarsh communities, habitats often present in coastal national parks. The impact of human trampling of benthic invertebrate habitat (pneumatophores and associated algae) and associated changes in gastropod communities at the landward – mid-region of a temperate mangrove forest (the area of highest gastropod diversity) were observed to be substantial, even at the equivalent of 25 people walking through a 30 cm wide undisturbed area (Ross, 2006). Hardiman and Burgin (2011b) undertook a similar study to Ross (2006) in a very different environment. In a pristine canyon environment of the Greater Blue Mountains World Heritage Area with very sandy substrate and low nutrient waters, they found that even at much higher levels of trampling than Ross (2006) had used, abundance and diversity of aquatic invertebrates returned to pre-trampling levels within one week. The impact of mountain biking may therefore differ greatly, even between aquatic ecosystems.

Appropriateness of national parks for mountain biking

While there is a dearth of information on the impact of mountain biking trails on the movement of fauna, Donaldson and Bennett (2004) reviewed the implications for internal fragmentation of parks and reserves due to roads and associated traffic. They concluded that the major ecological impacts were habitat alteration; constriction of the paths of animal movement; barriers to the movement of fauna; potentially isolating populations and communities; collision; and a source of biotic and abiotic effects. It was their view that, 'often to a lesser degree' these impacts were equivalent for recreational tracks used by bushwalkers. Based on the findings of a number of abovementioned studies (see Wilson and Seney 1994; Thurston and Reader 2001; Chiu and Kriwoken 2003) mountain biking would appear to have less impact than motorised vehicles. However, in contrast to most other forms of recreation that use access roads and paths in national parks, mountain biking trails are likely to be much more extensive and, at least over steeper sections, situated in much more vulnerable areas for the integrity of the local soils (i.e., steep, downhill slopes). With its demonstrated increasing popularity, the volume of bikers will exacerbate such issues and undoubtedly place more pressure on local fauna and their habitats.

Impacts associated with new forms of recreation have traditionally been handled by land managers by establishing normative standards for the activity and then developing rules and regulations to manage it (Ewert *et al.* 2006). Models such as 'Limits of Acceptable Change' have been developed to support such legislative decisions (Stankey *et al.* 1985), but may be difficult to apply to mountain biking if there are fundamentally differing perceptions of what constitutes acceptable use of public areas among stakeholders with strongly-differing views. In matters of such conflict it is often difficult for land managers to maintain their position. As a consequence, a typical process associated with the introduction of new recreational activities into a (protected) natural area is (1) resistance, (2) conflict, (3) compromise, and finally (4) accommodation (Ewert *et al.* 2006). This has apparently been the experience with mountain biking in UK national parks. The issues tend also to be exacerbated because visitors and managers perceive impacts differently (e.g. Martin *et al.* 1989; Hardiman and Burgin 2010b).

The power of advocacy through formal groups (e.g., sport specific associations) and informal (e.g., weblogs, online fora, social networking media) are becoming more influential and sophisticated in determining the appropriateness of national parks for mountain biking. The reach of the internet among members of increasingly technologically-knowledgeable user groups further enhances their lobbying power. Such networking has resulted in changes in decisions. For example, in the 1990s, in three UK national parks such campaigns resulted in changes in the decisions of land managers. Mountain biking in Dartmoor was initially made a criminal offence, while Exmoor considered it an 'unsuitable activity' and Snowdonia attempted to ban mountain bikers from its bridleways. Public opinion campaigns have since forced mountain biking to become an accepted activity in these parks, despite continued opposition (Palmer 2006).

Natural area managers are increasingly confronted with threats of appeals and litigation against their efforts to restrict what they perceive to be inappropriate recreational activities within protected areas. Such conflict with recreationists can be supported by the associated tourism and retailing industries who have commercial interests in the use of areas for their sport (Sarre 1989). The trail ahead therefore requires serious consideration of how best to deal with mountain biking and the associated degradation of natural areas, and potential loss of plant and animal species.

Lessons for balancing the rising demand for mountain biking in national parks against nature conservation, could be gleaned from the long-term conflict over horse riding. In 1996, a national park's survey in New South Wales revealed that 60% of park visitors objected to recreational horse riding in parks (Ramsay 1996) despite it being viewed as an integral part of the Australian image (Beeton 1999). The conflict continues (Newsome *et al.* 2002). We suggest that this may well be the outcome for mountain biking if the issues are not addressed adequately. To minimise the potential for such protracted conflict over mountain biking (or indeed any other emerging sport), decisions must be based on:

1. Sound ecological and social research;
2. Park management should genuinely engage with stakeholders (e.g., mountain bikers, other recreation users, relevant commercial interests, local residents, researchers) to develop options (these may include alternative venues to national parks);
3. Decisions, and clearly enunciated reasons that underpin these decisions should be widely disseminated; and
4. Monitoring activities, including studies to determine long term chronic impacts and on-going community

attitudes on an on-going basis to ensure that changes in management decisions are underpinned by defensible research.

Without a strong strategic approach to mountain biking that includes community engagement, underpinned by quality ecological and social science, the outcome will be further degradation of natural areas and, at the least, loss of many animals if not major threats to populations. We also predict that there will be on-going conflict between mountain bikers and other recreationists and residents.

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